High-resolution hydraulic parameter maps for surface soils in tropical South America


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Received: 28 September 2013 – Published in Geosci. Model Dev. Discuss.: 17 December 2013
Revised: 3 March 2014 – Accepted: 19 March 2014 – Published: 6 May 2014

Abstract. Modern land surface model simulations capture soil profile water movement through the use of soil hydraulics sub-models, but good hydraulic parameterisations are often lacking, especially in the tropics. We present much-improved gridded data sets of hydraulic parameters for surface soil for the critical area of tropical South America, describing soil profile water movement across the region to 30 cm depth. Optimal hydraulic parameter values are given for the Brooks and Corey, Campbell, van Genuchten–Mualem and van Genuchten–Burdine soil hydraulic models, which are widely used hydraulic sub-models in land surface models. This has been possible through interpolating soil measurements from several sources through the SOTERLAC soil and terrain data base and using the most recent pedo-transfer functions (PTFs) derived for South American soils. All soil parameter data layers are provided at 15 arcsec resolution and available for download, this being 20x higher resolution than the best comparable parameter maps available to date. Specific examples are given of the use of PTFs and the importance highlighted of using PTFs that have been locally parameterised and that are not just based on soil texture. We discuss current developments in soil hydraulic modelling and how high-resolution parameter maps such as these can improve the simulation of vegetation development and productivity in land surface models.

1 Introduction

Ecosystem water cycles are fundamental to our understanding of how vegetation develops, and how plants respond to periods of high and low water availability. Plants in all ecosystems obtain most of their water through soil, and the study of water movement through the soil matrix has a long history in both ecology and agriculture (Childs, 1969; Marshall et al., 1996; Leigh, 1999). Soil properties in general are widely recognised as one of the critical factors controlling ecological differences between and within biomes (Phillips et al., 1994, 2004; Leigh, 1999; Aragão et al., 2009; Lloyd et al., 2009; Quesada et al., 2012), as are soil hydraulic properties in particular (Marshall et al., 1996; Tomasella and Hodnett, 1998; Tomasella et al., 2000; Hodnett and Tomasella, 2002).

Soil information is a crucial input for vegetation and land surface modelling applications: soil properties strongly influence water exchange between the land surface and atmosphere as well as growth processes such as regeneration (e.g. Leigh, 1999; Marthews et al., 2008; Dharssi et al., 2009; Dadson et al., 2011). From a modelling perspective, soil water movement is the result of several overlapping processes, usually simulated by distinct hydrological sub-models (Dadson et al., 2011). At the particle scale within individual soil samples we find that capillary processes dominate (Townend et al., 2001; Hodnett and Tomasella, 2002; Fredlund et al., 2012); at the soil profile or site scale...
considerations of bypass flow arise (Marshall et al., 1996; Mathews et al., 2008), at the landscape scale relative topographic position and groundwater flow are the overriding factors (Dadson et al., 2011) and at the regional scale river routing and other transport processes can overwhelm all other effects (Dadson and Bell, 2010; Dadson et al., 2011). These different processes are all represented within land surface models, although with varying levels of sophistication (e.g. in the JULES model, Best et al., 2011).

For landscape-scale gridded model runs, pre-calculated ancillary files are required to provide the spatially varying parameter estimates required by hydrological sub-models (e.g. Dharssi et al., 2009; Castanho et al., 2013). The most widely used, publicly available set of global ancillary files are currently the IGBP-DIS parameter maps (Global Soil Data Task Group, 2000), however at 5 arcmin resolution (approximately 10 km at the Equator), these maps are now considered fairly coarse (Ke et al., 2012). Land surface models are now being applied at increasingly high spatial resolution both offline as well as coupled to climate models (Ke et al., 2012), which is necessary to capture the fine-scale dynamics of ecosystem development for realistic modelling of ecosystem productivity and development (Malhi and Wright, 2004; Mathews et al., 2008).

Tropical South America is the most intensively studied tropical region (Malhi and Wright, 2004) so we have taken this region as our focus. Across the tropical zone, the availability of soil-related data for running high-detail simulations has historically been low (Tomasella and Hodnett, 1998; Tomasella et al., 2000), but in recent years this zone has been recognised as one of the critical “driver” biomes of the world’s climate system and the situation is fast improving (Leigh, 1999; Malhi and Wright, 2004; Phillips et al., 2004). Much progress has been made in the availability of high-quality soil information (e.g. the SORi and TERrain data base SOTER, Dijkshoorn et al., 2005) and it is now possible to construct ancillary files of much higher resolution and reliability. Additionally, with ever more research groups gaining the capacity to carry out large-scale gridded simulations on a routine basis, there is an increasing need for spatially explicit parameter maps with which to drive those simulations (e.g. Castanho et al., 2013).

Soil hydraulic models are the components of land surface models that control and simulate water movement through the soil matrix (Table 1). Parameters for fine-scale soil hydraulic models are generally derived from soil profile data using pedotransfer functions (PTFs), which are equations used to estimate unavailable soil variables from closely related and more available soil properties such as texture and dry bulk density (Table 2). There is much current debate on the suitability of hydraulic models and PTFs in the tropics and we make recommendations for their use in the context of tropical South America (Fig. 1). Our hydraulic parameters may therefore be used to produce much more robust simulations of soil water dynamics for this region. Downloadable “model-ready” data grids of parameter values are produced from this analysis, in appropriate formats for use in a wide variety of land surface modelling applications. Finally, we discuss current developments in soil hydraulic modelling in general and identify ways in which these models can benefit from higher-resolution parameter maps such as these.

2 Methods

2.1 Soil hydraulic models

Water movement through saturated soil is described by Darcy’s law, which holds that the vertical rate of water movement is the product of a gradient in hydraulic potential and the soil hydraulic conductivity $k$ (Childs, 1969; Dullien, 1992; Marshall et al., 1996). In unsaturated soil Darcy’s law may be generalised to the Richards equation (Marshall et al., 1996), which involves $k$ becoming a function of the soil matric potential $\psi$ (aka matric tension, equivalent to capillary pressure head), where $\psi$ is the component of hydraulic potential when any differences in elevation are ignored (Marshall et al., 1996; Mullins, 2001). The relationships between $\psi$, $k$ and volumetric soil water content $\theta$ (unfrozen) can be described by two closely related curves called the soil water characteristic (SWC) and the soil hydraulic conductivity curve (HCC) (Fig. 2), which together describe the hydraulic model of the soil. In this study we use the four most widely used soil hydraulic models (Table 1).

2.2 South American soils

We focus on the soils of tropical South America (taken as the area from Panama to the tropic line from Antofagasta, Chile,
The four most widely used soil hydraulic models, expressed in terms of soil matric potential $\psi$ (Pa), volumetric soil water content $\theta$ ($\text{cm}^3 \text{cm}^{-3}$) and soil hydraulic conductivity $k$ ($\text{mm s}^{-1}$, vertical infiltration rate of water into the soil; $k_e = k/k_{sat}$ is the relative hydraulic conductivity). The van Genuchten–Mualem model is currently the de facto standard soil hydraulic model (Vereecken et al., 2010; also see comments in Dirksen, 2001): for reviews of these and other similar models see Shao and Irmaknejad (1999), Townend et al. (2001), Sillers and Fredlund (2001), Seki (2007), Pertassek et al. (2011) and Fredlund et al. (2012). See Pertassek et al. (2011) and Valiantzas (2011) for the combination Brooks and Corey–Burdine not included here. Matric potential $\psi$ is related to the capillary pressure head of soil water ($h$ in m) by Pascal’s law $h = -\psi/\rho g$ where $\rho = \text{volumic mass of water}$ ($1000 \text{ kg m}^{-3}$) and $g = \text{acceleration of gravity}$ ($9.8 \text{ m s}^{-2}$). Other standard quantities: $\theta_{sat}$ and $k_{sat}$ are the values of $\theta$ and $k$ when the soil is saturated (i.e. all soil pores are water-filled), $\theta_{res}$ is the residual soil water content (the soil water content never available to plants, taken here to be the value of $\theta$ when the soil is dry to $\psi = -16 \text{ MPa}$ or $1632 \text{ m}$ pressure head) and $\Theta = (\theta - \theta_{res})/(\theta_{sat} - \theta_{res})$ is the normalised volumetric water content (aka effective saturation $S_e$).

Table 1. The four most widely used soil hydraulic models, expressed in terms of soil matric potential $\psi$ (Pa), volumetric soil water content $\theta$ ($\text{cm}^3 \text{cm}^{-3}$) and soil hydraulic conductivity $k$ ($\text{mm s}^{-1}$, vertical infiltration rate of water into the soil; $k_e = k/k_{sat}$ is the relative hydraulic conductivity). The van Genuchten–Mualem model is currently the de facto standard soil hydraulic model (Vereecken et al., 2010; also see comments in Dirksen, 2001): for reviews of these and other similar models see Shao and Irmaknejad (1999), Townend et al. (2001), Sillers and Fredlund (2001), Seki (2007), Pertassek et al. (2011) and Fredlund et al. (2012). See Pertassek et al. (2011) and Valiantzas (2011) for the combination Brooks and Corey–Burdine not included here. Matric potential $\psi$ is related to the capillary pressure head of soil water ($h$ in m) by Pascal’s law $h = -\psi/\rho g$ where $\rho = \text{volumic mass of water}$ ($1000 \text{ kg m}^{-3}$) and $g = \text{acceleration of gravity}$ ($9.8 \text{ m s}^{-2}$). Other standard quantities: $\theta_{sat}$ and $k_{sat}$ are the values of $\theta$ and $k$ when the soil is saturated (i.e. all soil pores are water-filled), $\theta_{res}$ is the residual soil water content (the soil water content never available to plants, taken here to be the value of $\theta$ when the soil is dry to $\psi = -16 \text{ MPa}$ or $1632 \text{ m}$ pressure head) and $\Theta = (\theta - \theta_{res})/(\theta_{sat} - \theta_{res})$ is the normalised volumetric water content (aka effective saturation $S_e$).

<table>
<thead>
<tr>
<th>Model</th>
<th>Soil water characteristic (SWC)</th>
<th>Hydraulic conductivity curve (HCC)</th>
<th>Parameters</th>
</tr>
</thead>
<tbody>
<tr>
<td>Brooks and Corey–Mualem (B&amp;C–M)</td>
<td>$\theta = \left{ \begin{array}{ll} \frac{1}{\theta_{sat}} \forall \psi \geq \psi_e &amp; \text{for } \psi \geq \psi_e \ \frac{1}{\theta_{res}} \forall \psi &lt; \psi_e &amp; \text{for } \psi &lt; \psi_e \end{array} \right.$</td>
<td>$k_e = \left{ \begin{array}{ll} \frac{1}{\theta_{sat}} \left( L + \frac{2}{m} \right) &amp; \text{for } \psi \geq \psi_e \ \frac{1}{\theta_{res}} \left( L + \frac{2}{m} \right) &amp; \text{for } \psi &lt; \psi_e \end{array} \right.$</td>
<td>– Pore-size distribution index $\lambda (\geq 0, \text{ no units})$</td>
</tr>
<tr>
<td>Campbell (~ Mualem)</td>
<td>$\theta = \frac{e}{\theta_{sat}} \left{ \begin{array}{ll} \frac{1}{\theta_{sat}} \forall \psi \geq \psi_e &amp; \text{for } \psi \geq \psi_e \ \frac{1}{\theta_{res}} \forall \psi &lt; \psi_e &amp; \text{for } \psi &lt; \psi_e \end{array} \right.$</td>
<td>$k_e = \left{ \begin{array}{ll} \frac{1}{\theta_{sat}} \left( L + \frac{2}{m} \right) &amp; \text{for } \psi \geq \psi_e \ \frac{1}{\theta_{res}} \left( L + \frac{2}{m} \right) &amp; \text{for } \psi &lt; \psi_e \end{array} \right.$</td>
<td>– (Inverted) pore-size distribution index $b (\geq 0, \text{ no units})$</td>
</tr>
<tr>
<td>van Genuchten–Mualem (vanG–M)</td>
<td>$\theta = (1 + A)^{-mM}$</td>
<td>$k_e = \Theta \left( 1 - \left( 1 - \Theta \frac{m}{M} \right) \right)^{2} = (1 + A)^{-LM} \left( 1 - \left( \Theta \frac{m}{M} \right) \right)^{2}$</td>
<td>– Pore-size distribution index $n_M (\geq 1, \text{ no units})$</td>
</tr>
<tr>
<td>van Genuchten–Burdine (vanG–B)</td>
<td>$\theta = (1 + A)^{-mB}$</td>
<td>$k_e = \Theta \left( 1 - \left( 1 - \Theta \frac{m}{B} \right) \right)^{2} = (1 + A)^{-2mB} \left( 1 - \left( \Theta \frac{m}{B} \right) \right)^{2}$</td>
<td>– Pore-size distribution index $n_B (\geq 2, \text{ no units})$</td>
</tr>
</tbody>
</table>

* In any particular soil layer, $\theta_{res}$ is usually 5–30% smaller than the total porosity because of entrapped or dissolved air (van Genuchten et al., 1991). * This residual matric potential $\psi = -16 \text{ MPa}$ is approximately the limit of field conditions for desert soils (see discussion in Durner, 1994 on how to measure $\theta_{res}$). * $L$ may be restricted to $\geq 0.02$ on physical grounds (e.g. Matthews et al., 2008). * Correspondence between B&C–M and vanG–M: If you have values for the B&C–M parameters $\lambda$ and $\psi_e$ (assuming $L \equiv 1$) then Morel-Seytoux et al. (1996) proposed the following to calculate equivalent vanG–M parameters $n_M$ and $m_M$: $n_M = \lambda$ and $m_M = \psi_e / (\rho g)$. Derived from B&C–M by taking $\theta_{res} = 0 \text{ cm}^3 \text{cm}^{-3}, b = 1/\lambda$, and $L = 1$ (Campbell, 1974, 1985).
to São Paulo, Brazil, excluding the Galápagos Archipelago, and the outlying islands Cocos Is., Malpelo Is., Fernando de Noronha, St Peter and St Paul, Trindade and Martim Vaz and Rocas Atoll, Fig. 1). Tropical soils cover a huge range of types and this area is no exception (Ashton, 2004), with parent material varying from the Precambrian rocks of the Guiana and Brazilian Shields to the much younger Cenozoic geology of the Andes and western Amazon (Quesada et al., 2011).

Soil profile measurements were collected from three sources: (i) Quesada et al. (2010) collected data throughout the area as part of the projects RAINFOR http://www.rainfor.org/ and TROBIT www.geog.leeds.ac.uk/groups/trobit/, (ii) data from the large RADAMBRASIL project (de Negreiros et al., 2009) as supplemented by Cooper et al. (2005) provided good coverage of Brazil and (iii) data from the International Soil Reference and Information Centre (ISRIC) was used to cover the remaining area (Batjes, 2000). We focused on surface soil for this analysis because there was insufficient data for adequate mapping of deeper soil: all soil profile data from below 30 cm depth were discarded from these sources, which left a total data base of 7620 profile measurements across tropical South America. The soil areas of SOTERLAC (Dijkshoorn et al., 2005) were used, supplemented by data for the same polygons from Batjes (2010).

Over these profile measurements, values for cation exchange capacity (CEC) were used if available from Quesada et al. (2010) and the corresponding polygon in Batjes (2010) (68.8 % of points), otherwise values 10.1 cmol kg$^{-1}$ and 45.7 cmol kg$^{-1}$ were substituted for low and high activity clays, respectively (Hodnett and Tomasella, 2002), following the low- and high-activity categories in Batjes (2010) (Fig. 3). Values for soil organic carbon (SOC) were taken from RADAMBRASIL, Cooper et al. (2005) and Quesada et al. (2010). Values for dry bulk density (DBD) were used if possible from Batjes (2000) (1.2 % of points), otherwise a value of max(0.05, 1.578 − (0.054 × (SOC/10)) − (0.006 × SIPC) − (0.004 × CLPC)) g cm$^{-3}$ was substituted (Tomasella and Hodnett, 1998) (a further 90.0 % of points; all variables and units are given in Table 2).

2.3 Spatial analyses

All GIS data analyses were carried out in ArcGIS 10.0 (Esri Inc., Redlands, CA). Taking the SOTERLAC polygons as the base areal units, mean values of available soil properties were calculated and assigned to the polygon containing those points. If no profile measurements were available for a particular polygon, a taxotransfer rule was followed with a mean assigned from all profile measurements in the same soil type (taken from Dijkshoorn et al., 2005). In the few cases where no measurements were available from a particular soil
Table 2. Three sets of PTFs, of which the Hodnett and Tomassella (2002) PTFs are the ones we recommend for general use in tropical South America (for reviews of PTFs see Tomassella and Hodnett, 1998, 2004; Wösten et al., 1999; Tomassella et al., 2000, 2003; McBratney et al., 2002; Vereeken et al., 2010; Rasoulzadeh, 2011; Pan et al., 2012). Independent variables: CLPC = clay fraction (%), SAPC = sand fraction (%), SIPC = silt fraction = 100 − CLPC − SAPC, DBD = dry bulk density (Mg m⁻³ or, equivalently, g cm⁻³), CEC = cation exchange capacity (cmol kg⁻¹), SOC = soil organic carbon content (g C kg⁻¹ soil), pH = hydrogen ion activity. For Brooks and Corey and van Genuchten, there is some evidence that L varies between soil categories (e.g. Wösten et al., 1999; Schaap and Leij, 2000; ECMWF, 2009, Table 7.6), but we are not aware of any continuous, tropically based PTF for this parameter and suggest simply to assume L = 0.5. PTFs should be applied only with caution outside the geomorphic region or soil type from which they were developed (McBratney et al., 2002). Additionally, the relationship between soil pore-size and particle-size distributions is only approximate, especially in highly leached soils and low-acidity classes: SWC curves are closely related to the pore-size distribution, whereas soil texture is related to particle sizes, so there is inherent uncertainty involved in using PTFs based only on texture (Vereeken et al., 2010).

Source | PTFs (see Table 1 for quantity definitions) |
-------|------------------------------------------------|
Cosby et al. (1984)⁴ | (£, dimensionless) = 1/((1.04 + 0.1675 × CLPC − 0.003 × SAPC) |
Non-tropical and texture-based, but widely used | (b, dimensionless) = 1/3.11 + 0.1575 × CLPC − 0.003 × SAPC |
(ψₘ in Pa) = −0.01 (10 × (2.17 − 0.0015 × CLPC − 0.0158 × SAPC)) × (1000 − 9.8065) |
(ψₚ in m of pressure head, also called kₑ) = −(ψₘ in Pa)/(μ/ρ) = 0.01 (10 × (2.17 − 0.0063 × CLPC − 0.0158 × SAPC)) |
(βₑ₀ in cm³ cm⁻³) = 0.01 × (50.5 − 0.037 × CLPC − 0.142 × SAPC) |
(kₑ₀ in µm s⁻¹ or, equivalently, kg m⁻³ s⁻¹) = (25.4/60.60 × 10−3 × (1 − 0.60/0.0064 × CLPC + 0.0126 × SAPC)) |
Cosby et al. (1984) estimated the uncertainty of log₁₀(ψₑ₀/(25.4/60.60)) to be < 0.7, i.e. the uncertainty of kₑ₀ is < 10⁻¹·⁵¹₄⁷ = 0.0354 mm s⁻¹ (a wide bound, but consistent with theory discussed in Dirksen, 2001). |
Tomasella and Hodnett (1998) | (£ₜ, dimensionless) = exp(−((1.197 + (0.00147 × SIPC) − (0.0445 × CLPC) + (0.000894 × SIPC + CLPC) − (0.0001 × SIPC + CLPC))) |
Tropical, but still texture-based | (b, dimensionless) = 1/3.11 + 0.1575 × CLPC − 0.003 × SAPC |
(ψₘ in m of pressure head, also called kₑ) = −(ψₚ in Pa)/(μ/ρ) = 0.01 (10 × (2.17 − 0.0037 × CLPC − 0.0158 × SAPC)) |
(βₑ₀ in cm³ cm⁻³) = 0.01 × (40.61 + (0.165 × SAPC) + (0.162 × CLPC) + (0.0137 × SIPC + CLPC) + (0.00018 × SIPC + CLPC)) |
(kₑ₀ in µm s⁻¹ or, equivalently, kg m⁻³ s⁻¹) = 0.01 × max(0, −2.094 + (0.047 × SOPC) + (0.431 × CLPC − 0.00827 × SIPC)) |
Hodnett and Tomassella (2002, Table 8), Tomasella and Hodnett (2004) | (£ₜ, dimensionless) = exp((62.986 − (0.833 × CLPC) − (0.529 × (SOC/10)) + (0.593 × (CLPC + CLPC) − (0.007 × CLPC + CLPC) − (0.014 × SIPC + CLPC))/100) |
Tropical and more sophisticated than texture-based PTFs | (a in m⁻¹) = (1000 × 9.8065)/100 × (1 − 0.7652 × SIPC + (2.440 × (SOC/10)) − (0.076 × CECl − (11.331 + ph) + (0.019 × SIPC + CLPC)))/100) |
(α in m⁻¹) = (a in m⁻¹)/(μ/ρ) = 0.01 × exp((−2.294 − (3.526 × SIPC) + (2.440 × (SOC/10)) − (0.076 × CECl − (11.331 + ph) + (0.019 × SIPC + CLPC)))/100) |
(βₑ₀ in cm³ cm⁻³) = 0.01 × (81.799 + (0.099 × CLPC) − (31.42 × DBD) + (0.018 × CECl) + (0.451 × ph) − (0.0005 × SAPC + CLPC))/100) |
(kₑ₀ in µm s⁻¹ or, equivalently, kg m⁻³ s⁻¹) = 0.01 × (22.733 − (0.164 × SAPC) + (0.235 × CECl) − (0.831 × ph) + (0.0018 × CLPC + CLPC) + (0.0026 × SIPC + CLPC))/100) |
Hodnett and Tomassella (2002) estimated the uncertainty of £ₜₓ (a in m⁻¹), £ₑₐₜ and £ₑₘ in ±0.4382, 5.583 m⁻¹², 0.1007 cm³ m⁻³ and 0.0765 cm³ m⁻³, respectively (mean uncertainties across texture classes). |

Example calculation using values from Tambaopora forest plot TAM-05, Peru (12°10′49′′ S, 69°16′14′′ W, Phillips et al., 2004). At this location, the PTFs from Hodnett and Tomassella (2002) give the following values: £ₜₓ = 1.08 ± 0.43 m⁻¹, £ₑₐ = 7.62 ± 0.58 m⁻¹², £ₑₘ (from Cosby et al., 1984) = 0.0024 mm s⁻¹ (uncertainty ±0.0354, but must be > 0.66 mm s⁻¹), £ₑₚ = 0.86 ± 0.10 cm³ m⁻³ and £ₑₖ = 0.25 ± 0.08 cm³ m⁻³ (high values because the soil is a clay and the six soil profile measurements for Tambaopora in our data base all had high SOC (mean = 501 g kg⁻¹) and low CEC (mean = 4.03 mol kg⁻¹), indicating likely microaggregation). |

Put in ψ = −150 000 Pa for permanent wilting point (see Table 1 for equations, taking L = 0.5): h = −ψ/(1000 × 9.8065) = 152.9 m, A = (aₑₚ) = 2117.3, θ = (1 + A) × (1/1-1) = 0.55 so £ₑₚ = (£ₑₜ - £ₑₚ) + £ₑₚ = 0.58 cm³ m⁻³ |

Put in £ₑₚ = −100 000 Pa for field capacity (we use the standard −10 kPa not −33 kPa, following Marshall et al., 2002; Townsend et al., 2001; Tomasella and Hodnett, 2004): h = 1.02 m, A = 9.2, θ = 0.83 so £ₑₚ = 0.76 cm³ m⁻³ |

so the Available Water Capacity of this soil down to 30 cm depth is $\text{AWC} = (\psi_{wp} - \psi_{up}) = 0.30 \times (0.76 - 0.58) = 0.054 m^3$ water or 54 mm water (g. m. E. McBratney et al., 2002). The Profile Available Water Capacity of this soil down to the 1.7 m depth of the soil column (Tomasnell et al., 2001), assuming that this soil is shallow enough for these surface parameters to be assumed constant over the soil column, is $\text{PACW} = (\psi_{EC} - \psi_{wp}) \times depth = (0.76 - 0.58) \times 1.7 = 310$ mm water or 180 mm water m⁻¹ soil over the soil column (slightly lower than might be expected given the high £ₑₚ value).

* Using the standard relationship £ₑₚ = 1/((1.04 + 0.1675 × CLPC − 0.003 × SAPC) DBD in g cm⁻³ and soil particle density £ₑₚ in g cm⁻³ (Marshall et al., 1996) and an assumed approximation for £ₑₚ (e.g. £ₑₚ = 1.25 × SOM + 2.65 × (100 − SO4)/100 g cm⁻³ where SOM = (SOC/10)0.58) it is possible to replace the given PTFs for £ₑₚ, however uncertainty in the value of £ₑₚ is generally high at continental scales so this approach was not used here. In theBrowse land surface model these equations were previously misquoted with exponentially exp) rather than linear units of m 10⁻¹ (Kowalchuk 2008; Biber et al., 2011) * For £ₑₚ many sources use units £ₑₚ in kg ms⁻¹ or £ₑₚ in m⁻¹ which is equivalent to each other (see Marshall et al. 1996, p. 37) but a factor of 10 different from these units.
Climate and Forecast (CF) conventions. Considering the soil quantities from Table 2 in order:

- The saturated hydraulic conductivity $k_{\text{sat}}$ is lower across the Amazon Basin than it is across the Cerrado (Fig. 4a), with mean across the domain $k_{\text{sat}} = 0.008 \text{ mm s}^{-1}$ ($n = 1902$ SOTERLAC polygons), which is comparable to the $k_{\text{sat}} = 0.010 \text{ mm s}^{-1}$ typical value for a microaggregated clay soil under tropical forest in Panama (Marthews et al., 2008). A reminder: this is a value for surface soil < 30 cm depth (as are all other values given here) and $k_{\text{sat}}$ usually decreases with depth (Elsenbeer et al., 1999; Clark and Gedney, 2008).

From the Tomasella and Hodnett (1998) PTFs:

- Parameter $b$ is broadly a measure of the steepness of the SWC (freely draining sandy soils generally have low values of $b$ (high values of $\lambda$) whereas heavy clay soils have high $b$ values). Values derived for tropical South America are uniformly lower (mean $= 4.8$) than those derived from applying the Cosby et al. (1984) PTF for $b$ (mean $= 6.9$).

- The air-entry potential $\psi_e$ is the matric potential $\psi$ at which the soil first desaturates when drying after heavy rainfall (i.e. at which the largest pores drain). Values derived for tropical South America are less negative (mean $= -1.1 \text{ kPa}$) than those derived from applying the Cosby et al. (1984) PTF for $\psi_e$ (mean $= -2.5 \text{ kPa}$).

- The residual soil water content $\theta_{\text{res}}$ values are generally much lower (mean $= 0.04 \text{ cm}^3 \text{ cm}^{-3}$) than the $0.1866 \text{ cm}^3 \text{ cm}^{-3}$ value suggested by Tomasella et al. (2000, Table 3) for Brazilian soils.

- The saturated soil water content $\theta_{\text{sat}}$ values are generally higher (mean $= 0.51 \text{ cm}^3 \text{ cm}^{-3}$) than those derived from applying the corresponding Cosby et al. (1984) PTF (mean $= 0.43 \text{ cm}^3 \text{ cm}^{-3}$) and closely match the $0.502 \text{ cm}^3 \text{ cm}^{-3}$ value suggested by Tomasella et al. (2000, Table 3) for Brazilian soils.

- Using the parameter correspondences of Morel-Seytoux et al. (1996, Table 1), estimated values for the van Genuchten–Mualem model may also be calculated from these PTFs, giving mean $n_M = 1.21 \alpha = 0.53 \text{ m}^{-1}$.

From the Hodnett and Tomasella (2002) PTFs:

- The van Genuchten parameter $n_M$ is broadly a measure of how uniform are pore sizes in the soil and this gives $n_M$ values slightly lower (mean $= 1.47$, Fig. 4b) than the 1.571 value suggested by Tomasella et al. (2000, Table 3) for Brazilian soils. In deeper soil layers, $n_M$ should increase (assuming SOC decreases with depth in the PTF of Hodnett and Tomasella, 2002).
Fig. 4. Soil hydraulic parameter maps for tropical South America derived from the PTFs of Cosby et al. (1984) (a) and Hodnett and Tomasella (2002) (b–e): (a) saturated hydraulic conductivity \( k_{\text{sat}} \) (mm/s), (b) Van Genuchten parameter \( n_M \), (c) Van Genuchten parameter \( \alpha \) (m\(^{-1}\)), (d) saturated soil water content \( \theta_{\text{sat}} \) (cm\(^3\)/cm\(^3\)) and (e) residual soil water content \( \theta_{\text{res}} \) (cm\(^3\)/cm\(^3\)). National borders and the outline of Amazonia sensu stricto are also shown (Eva et al., 2005). For example, the calculated values given for Tambopata (Table 2) were from an area in SE Peru with unusually high residual and saturated soil water content in its soils and maps (c) and (d) give an indication of where soils of similar hydraulic properties may be found elsewhere in the region.

- The van Genuchten parameter \( \alpha \) is broadly a measure of how structured the soil is and this gives \( \alpha \) values much lower (mean = 3.1 m\(^{-1}\), Fig. 4c) than the \( \rho_g \times (1.0631/1000) = 10.4 \) m\(^{-1}\) value suggested by Tomasella et al. (2000, Table 3) for Brazilian soils. In deeper soil layers, \( \alpha \) should decrease (assuming SOC decreases with depth in the PTF of Hodnett and Tomasella, 2002).

- The residual soil water content \( \theta_{\text{res}} \) values are much higher (mean = 0.19 cm\(^3\)/cm\(^3\), Fig. 4d) than those derived from applying the corresponding Tomasella and Hodnett (1998) PTF (mean = 0.04 cm\(^3\)/cm\(^3\)), closely matching the 0.1866 cm\(^3\)/cm\(^3\) value suggested by Tomasella et al. (2000, Table 3) for Brazilian soils. In deeper soil layers, \( \theta_{\text{res}} \) should be approximately the same as the surface value (neither DBD nor SOC are in the PTF of Hodnett and Tomasella, 2002).

- The saturated soil water content \( \theta_{\text{sat}} \) values are slightly lower (mean = 0.48 cm\(^3\)/cm\(^3\), Fig. 4e) than those derived from applying the corresponding Tomasella and Hodnett (1998) PTF (mean = 0.51 cm\(^3\)/cm\(^3\)) and slightly lower than the 0.502 cm\(^3\)/cm\(^3\) value suggested by Tomasella et al. (2000, Table 3) for Brazilian soils. In deeper soil layers, \( \theta_{\text{sat}} \) should decrease (assuming DBD increases with depth in the PTF of Hodnett and Tomasella, 2002).

Over tropical South America the Hodnett and Tomasella (2002) PTFs give values that appear to be a great improvement over those from Cosby et al. (1984) and Tomasella and Hodnett (1998) (mean SWC curves based on these for standard USDA soil textural classes are given in Hodnett and Tomasella (2002), if required). Therefore, we recommend these functions whenever data are available for their implementation.

4 Discussion

Our study domain is all of tropical South America, covering all or part of 13 countries. We have also covered the entirety of the Amazon Basin – the most intensely studied part of the tropics – and the whole of the Amazon forest biome (containing approximately 50% of global tropical forests and > 50% of all species that exist on Earth) as well as some 20% of the world’s freshwater resources (Malhi and Wright, 2004).
Tropical South America is a critically important region and methods shown to work well here will set the standard for many other areas of the world.

Land surface models are being applied at ever increasing spatial resolution as a means to model ecosystem productivity and development as realistically as possible (Malhi and Wright, 2004; Marthews et al., 2008; Ke et al., 2012). Characterising soil physical properties is a key element in modelling land surface–atmosphere exchange processes and, therefore, critical to the successful application of coupled land surface models. The van Genuchten–Mualem model is the current de facto standard soil hydraulic model (Vereecken et al., 2010) and the downloadable data sets provided by this paper are sufficient to parameterise this model across tropical South America, as well as to parameterise three widely implemented alternatives (Table 1). The maps provided here are of 20× higher resolution than the 5 arcmin resolution IGBP-DIS parameter maps, which are the best previously available (Global Soil Data Task Group, 2000; also see Ashton, 2012) and based on high-quality field data (Quesada et al., 2010) and the most comprehensive soil survey to date for South America (SOTERLAC, Dijkshoorn et al., 2005) and the latest PTFs from field data (Hodnett and Tomasella, 2002).

4.1 Improving model parameterisation

As with all models, soil hydraulics models can only be as good as their parameterisation from field data (see discussions in McBratney et al., 2002; Vereecken et al., 2010; Ke et al., 2012). By producing high-quality parameter maps based on pedotransfer functions (PTFs) that are (i) more sophisticated than simple texture-based PTFs (Table 2, specifically including dependence on cation exchange capacity (CEC) which can account at least partially for microaggregation effects) and were (ii) derived from local soil profiles in tropical South America (Hodnett and Tomasella, 2002), these parameter values are as robust as possible using currently available data sources. Additionally, the approach and tools we have developed are applicable, with the substitution of locally derived PTFs, in all other areas of the globe.

Uncertainties in all predicted quantities are given following the estimates quoted in our sources (see Table 2). Note that because many model simulations using these layers may be run at coarser resolution than 15 arcsec (e.g., climate models, usually run at resolutions of at least 0.5°), an aggregation step may be required to produce layers in a model-ready format: Aggregation may introduce uncertainties in addition to the base uncertainty of our layers (given in Table 2), dependent on the aggregation algorithm used.

4.2 The wider context: future model development

The soil hydraulic models presented in Table 1 are the current state of the art, and all can describe soil water movement well in many soils. However, they are not without limitations and improving these models is an active area of research (e.g., Vereecken et al., 2010; Fredlund et al., 2012). The value of the detailed parameter maps presented in this study would be greatly enhanced if progress could be made on four widely recognised model improvements in particular:

Improved modelling of macropore flow. Macropore flow is dictated by soil macrostructure rather than microstructure or texture and there is a recognised deficit of knowledge in this area (Lin et al., 2010). For example, the assumption of a unimodal pore-size distribution in standard soil hydraulics models is a significant barrier to further progress (Durner, 1994; Vereecken et al., 2010), especially in areas where microaggregated soils are common (Hodnett and Tomasella, 2002). Modelling macropore flow is especially important for characterising nutrient retention and leaching effects and improving the model representation of real, bimodal soil pore-size distributions could greatly improve the simulation of soil water movement in general (see e.g., Durner, 1994; Kutilek, 2004; Schaap and van Genuchten, 2006; Russell, 2010; Schelle et al., 2010; Zeilinger et al., 2010; Vereecken et al., 2010; Lin et al., 2010). For example, models describing dual-porosity or double-porosity media in the context of flow through fractured rocks may be used to describe bypass and capillary flow in soil columns (Gerke and van Genuchten, 1993; Adler et al., 2005; Guarracino and Monachesi, 2010). In the case where it is unfeasible to move to a multimodal pore-size distribution model, the use of a parameterisation PTF that includes CEC would nevertheless be a significant improvement because that avoids grouping low- and high-activity clays together, resulting in a more realistic placing of the modal pore-size and therefore improved simulations of microaggregated soils where they occur.

Improved modelling of very dry soil. Soil that is exposed to direct sunlight and is within tens of cm of the surface can become much drier than permanent wilting point (PWP, Fig. 2) and this can occur frequently under natural conditions even in humid forests, for example during short dry spells in a wet season (Marthews et al., 2008). In very dry soil, the liquid water content is no longer continuous so capillary forces become negligible and the strength of adsorptive forces controls water retention (Vereecken et al., 2010; Fredlund et al., 2012). Standard soil hydraulics models do not model these effects well, simply assuming that a certain amount of water is “unavailable” to plants (Table 1) even though this approach has clear limitations (Durner, 1994; Vereecken et al., 2010). Improving the model representation of desorptive drying in very dry soil could greatly improve the simulation of soil water movement in general (see e.g., Sillers and Fredlund, 2001; Schelle et al., 2010; Vereecken et al., 2010; Fredlund et al., 2012). This is especially important in environments that are semi-arid or have a pronounced dry season.

Improved modelling of tortuosity. The pore-size distribution (PSD) describes the porosity of the soil (the proportion of micropores and macropores) but how they connect to each other through the soil is described by a pore
connectivity model (PCM) and the current almost-universal standard PCM remains that of Mualem (1976). Mualem’s PCM included tortuosity as an exponent (parameter \( L \)) and work in this area since then has mostly concentrated on deducing or fitting optimal values for \( L \) (see Table 1). However, it should be remembered that Mualem (1976)’s study of 45 soils only included two clays and three clay loams and no tropical soils at all. Despite its wide use, it seems premature to assume that the Mualem PCM’s functional form is optimal for all soil types, and we note that alternative PCMs do exist (see reviews in Dullien, 1992; Kutilek, 2004).

**Inclusion of deeper soil layers.** The general worldwide lack of soil hydraulic data from below approximately 1.5 m depth, especially regarding hydraulic conductivity, is currently an issue of great importance in soil science and remains a major impediment to the modelling of water movement and uptake in deep-rooted ecosystems such as tropical forests and savannas (Hodnett and Tomasella, 2002). In tropical South America, for example, there remains today an almost complete lack of soil profile data from deep soils (M. Hodnett, personal communication, 2013) because of the difficulty in extracting uncompacted and otherwise undisturbed soil cores (see discussion of field methods in Marshall et al., 1996; Dirksen, 2001, although some protocols do extend to deeper layers, e.g. RAINFOR-GEM soil sampling to 4 m, Mathews et al., 2012). In the absence of data from deeper layers, extrapolations are usually made based on topsoil properties, but this is uncertain because of the usually greater compaction of subsoils and the possible presence of markedly different soil layers (e.g. acidic or sodic layers, impermeable layers at shallow depth, Marshall et al., 1996; Lloyd et al., 2009). Some soil parameters are generally assumed to remain constant with depth (e.g. \( \lambda \)) and others assumed to decrease exponentially (e.g. \( k_{\text{sat}} \)). Clark and Gedney, 2008; although note that in many soils \( k_{\text{sat}} \) actually increases with depth, e.g. sапролитic layers or carbonate outcrops that have undergone karstification), but a full hydropedological survey is required to accurately describe the hydraulic properties of a soil profile at all depths.

In this study we have restricted our analysis to 30 cm soil depth because of the need for a consistent data base to extrapolate over a continental area, however understanding the soil sub-surface can nevertheless be crucial. For example, soils in the Acrisols group (Alisols, Luvisols, Lixisols and Acrisols), which cover \( \sim 30\% \) of Amazonia, have a typical two-layer vertical particle size distribution, being quite sandy in the top soil and clay rich in the subsoil (usually just below 30 cm). For such soils, the parameter estimates and model fits presented in this study apply only to the upper layers and the effects of the lower layers should be accounted for separately.

Despite their sophistication in above-ground aspects, land surface models struggle to simulate systems where sub-surface flow decouples deeper soil layers hydrologically from upper layers, thereby greatly affecting estimates of ecosystem resilience in the long term (e.g. see Elsenbeer et al., 1999 and discussions in Lloyd et al., 2009 and Dadson et al., 2011). A recent example of this is the “Australian Evergreen Paradox” where the marked dry season of the northern Australian monsoon tropics would lead a land surface model to predict dominance by deciduous tree species, while the actual dominance by evergreen species is a consequence of deep-water aquifers (Bowman and Prior, 2005).

### 4.3 Spatial patterns of ecosystem function

Soil properties and the dynamics of soil water movement are a fundamental control on the development and productivity of all ecosystems (Lloyd et al., 2009). e.g. Aragão et al. (2009) drew attention to edaphic controls on forest productivity in the Amazon Basin. However, there is a need to separate out which particular soil properties affect vegetation and forest structure in a region such as tropical South America (Quesada et al., 2010, 2011, 2012). Land surface models are an ideal tool for this task because they are process-based and can be used to isolate individual causes and effects, but the current generation of land surface models struggles to address this task at continental scales (e.g. Castanho et al., 2013). We believe that the problem may not be a lack of sophistication in these models, but – more simply – uncertainties involved in the parameter values used to calibrate them.

For example, it is known that sites in western Amazonia with younger, more fertile soils often have poorer physical properties than sites in central and eastern Amazonia with older, highly weathered soils (e.g. shallower soil depth, lower drainage capacity, Quesada et al., 2010, 2012) and this has been correlated with variation in floristic composition and tree turnover rates across the Amazon Basin (Phillips et al., 2004; Quesada et al., 2012). Similarly, in NE Amazonia generally tree wood is denser and seeds are larger, which has been correlated with the poorer soils that also occur there (ter Steege et al., 2006). However, it is not easy to separate cause and effect here: for example, younger soils are not always more fertile than older soils, and correlations between tree turnover rates and soil physical properties are always potentially confounded with uncontrollable factors such as restricted species ranges and carbon fertilisation rates (see discussions in Phillips et al., 2004 and Ashton, 2004). Some soil properties show a broad east–west variation (e.g. high vs. low-activity clays, Fig. 3), but other parameters present a more complex spatial pattern – notably soil hydraulic parameters, as shown in this paper (Fig. 4). Spatial patterns such as these imply complex spatial variation in forest dynamics that can only be described very superficially by the current generation of land surface models. Investigations of the causes of these little-studied spatial patterns requires high-resolution ancillary files exactly like those presented here.
5 Conclusions

With recent improvements in the availability of spatial soil data grids (e.g. SOTER, Dijkshoorn et al., 2005) and improved region-specific PTFs (Tomasella and Hodnett, 2004), it has become possible to generate regional parameter maps for all models of soil hydraulics at high resolution. In this paper we have produced gridded data sets for tropical South America that are a step improvement on the best comparable maps currently available to the public.

Variation in soil hydraulic parameters may explain less of the known variation in forest dynamics and ecosystem productivity across tropical South America than other factors such as species-specific responses (Leigh, 1999; ter Steege et al., 2006) or climate (e.g. Malhi and Wright, 2004; Lloyd et al., 2009). However, it does seem clear that we currently know far too little about the portion of that variation that is under the control of these parameters. We recognise a pressing need for improved understanding of the various processes that control soil water dynamics, especially soil structural and microstructural aspects of tropical soils. There is also a need for soil data sets that extend profile information to depths of at least 1.5 m in order both to include a greater proportion of the rooting zone and to determine where and when reduced saturated hydraulic conductivity causes lateral flow.

Land surface models are experiencing a time of rapid development, but in some ways code development has progressed more quickly than development of the parameterisations on which code simulations are based (e.g. Ke et al., 2012). The time is right for a strong improvement in the quality of parameterisation behind these models, which will lead to much more robust simulations of soil water dynamics and, ultimately, greatly improved vegetation and biome productivity predictions for the tropical zone as a whole.

Acknowledgements. This study is a product of the Andes Biodiversity and Ecosystem Research Group (ABERG, http://www.andesconservation.org/) and has drawn heavily on its collaborators, infrastructure and data sources. We are indebted to the Gordon and Betty Moore Foundation (grant to RAINFOR) and the Jackson Foundation and Oxford Martin School (grants to Y. Malhi). C. A. Quesada was supported in part by the European Union Fifth Framework Programme as part of CARBONSINK-LBA, part of the European contribution to the Large-Scale Biosphere–Atmosphere Experiment in Amazonia (LBA), and by the Gordon and Betty Moore Foundation. Thanks to J. Schmerler, M. Schwartz, G. Lloyd, G. Borges, E. Oblitas, J. Morais Filho, R. Araújo Filho and O. Ferreira Cruz Junior for their field and laboratory assistance. Thanks to S. Abele for help with Python code and mapping and P. Tempel, A. Zeiliguer, G. Hill and J. Lloyd for helpful discussions during development. Finally, huge thanks to I. Perzia for continued support in every way during the writing of this paper.

Edited by: O. Marti

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